

A Spatio-Temporal Distribution and Dynamics of Major-, Minor- and Trace- Elements in Two Temperate Urban Lacustrine Ecosystems of Srinagar, Kashmir

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Abstract. Water bodies act as indicators of the extent of anthropogenic interference and pressure while evolving exhibit fluxes in intrinsic homeostasis. The ecological assessment of standard biophysicochemical variables through space-time monitoring in the region of anthro-urban intensification identifies such ecosystem responses. Sampling, preservation, preparation and analysis follow typical procedures for all the representative testers. Statistical procedures for inference of results involve Mean, Standard Deviation, Range, Correlation, Regression, Analysis of Variance (ANOVA), Cluster Analysis and Principal Components Analysis (PCA). Temperature, pH, conductivity and ionic composition of the surface waters have $p < 0.01$ and $R^2 > 0.5$. Total P perimeter outdid critical eutrophic index ($\leq 0.05 \text{ mgL}^{-1}$) but nitrate N persisted underneath it ($\leq 0.5 \text{ mgL}^{-1}$) ordinarily. Elevated nutrient levels reflect multiple human origins and catchment interventions. Priority pollutants (As, Cd, Cr, Cu, Pb, Ni and Se) remained below chronic levels with exception of Fe and Zn exceeding maximum permissible limits for irrigation purpose of the water. Major elements in surface sediments are Si > Ca > Mg > K > Na > P > S > Cl besides the micro and trace ones follow as Fe > Al > Zn > Mn > Cu > Cr > Ni > Co > As > Sn > Pb > Cd. Sediments correspond to calcite catchment character to conform their sink-selves. The peak biomass values as gm^{-2} are 880.2 in case of *Myriophyllum aquaticum*, 678.4 in *Nelumbo nucifera*, 182.4 in *Ceratophyllum demersum* and 45 in *Salvinia natans*. The upgraded biomass configuration coincides with supplementary bioconcentration. C/N < 10 implies low decomposition rates and autochthonous origin of organic matter. A semi-drainage type of flushing hydrology ensures low basin holding time for self-reclaimed nutrient balance. Sediment dredging is beneficial for internal nutrient deloading only after external inputs are curtailed. Biological scavenging and biogenic calcification together lower the bioavailability of nutrients and toxics. Hence, the systemic insitu pliability can prevail till exsitu lacustrine management fortifies euoecism to evade detrimental dysoecism.

Keywords: Sediment chemistry, productivity, organic matter, trophic status, trace-elements.

1 Introduction

Lakes are natural or artificial, brackish or freshwater lentic impoundments located within inland geologic depressions of Earth's crust. Their ephemeral feature manifests in originating due to certain geological processes and terminating as a result of disrupted hydrological balance or sedimentation [1]. Lacustrine ecosystems possess a dynamic equilibrium among and between physical, chemical and biological attributes responding to natural and anthropogenic fluctuations. The transition from oligotrophy to eutrophy steered by nutrient enrichment, sedimentation and higher productivity eventually transforms a lake into dystrophic marsh [2]. The diversified anthropogenic activities have led to the global sprawl of spectra of contaminants intoxicating various environmental compartments and posing potential risks for biota and human health as well [3] and [4]. Consequently, the natural setting and homeostasis of aquatic ecosystems is continuously and significantly altered by progressive socioeconomic, urban, industrial and agricultural developments [5], [6] and [7]. The impacts of the human activities on the overall surface-water quality devalue its suitability for a designated purpose besides influencing the structure and function of its abiotic and biotic subsystems [8]. Both the eutrophication process and water quality classification in lakes is subject to quantum of principal chemical constituents and pollutants distributed within the water column, sediment and biota. Water quality [9], sediment characteristics [10], biological

indicators [11] and even key ecological processes like primary production are essential to assess the integrity and health of the ecosystem [12]. The temporal variation in ionic composition of natural waters along with its spatial distribution in a region aids significantly to distinguish the source [13]. It is incumbent to conduct continuous field monitoring studies to understand the characterisation and contribution of each contaminant responsible for environmental deterioration to evolve sustainable eco-conservation strategies, practices and technologies [14], [15] and [16]. Therefore, in order to understand the level of contamination and innate capacity of the lake components to dissipate it for wellbeing of the ecosystem is undertaken for finding provisions of prospective ecomanagement approaches.

2 Materials and Methods

2.1 Study Area

Water, sediment and macrophytes from 50 sampling stations of five study sites of Anchar and Dal lakes (Fig. 1) were collected from March, 2011 to February, 2013. The five sites designated as A1, A2, D1, D2 and D3 were each further fragmented into ten sampling stations in order to obtain composite representative samples in triplicate for enhanced precision. Due consideration is given to collect and preserve the sub-samples of water, sediment and hydrophytes from the same spot.

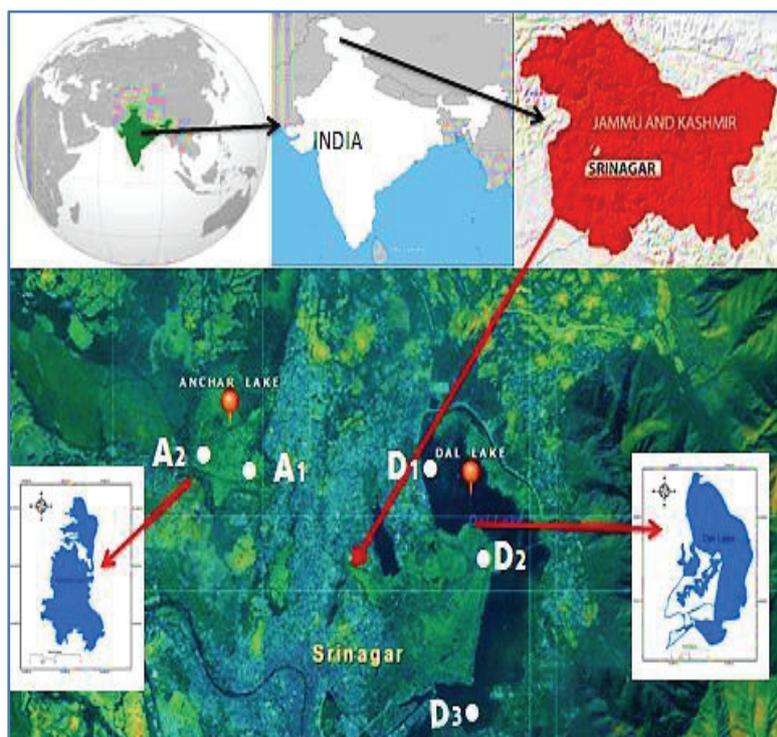


Figure 1. Location of study area.

2.2 Physico-chemical Analysis of Water

The water samples were analysed for temperature, pH, conductivity, bicarbonates, chloride, calcium, magnesium, sodium, potassium, nitrate nitrogen and total phosphorus on monthly basis while trace elements of Al, As, Cd, Co, Cr, Cu, Fe, Hg, Mn, Ni, Pb, Se, Sn and Zn were measured for spring, summer and winter seasons. The collection, preservation, preparation, storage and estimation of water samples follow standard methods [17], [18], [19], [20], [21] and [22].

2.3 Chemical Analysis of Sediments

The sediment analysis includes H-ion concentration, electrical conductivity, bicarbonates, organic-C, organic matter, total-N, C/N ratio, elemental composition of Ca, Cl, K, Mg, Na, P, S, Si and the given trace elements for each season. Electrical Conductivity is determined from filtrate of 1: 2 W/V sediment- distilled water suspension using a standard 0.01 M KCl solution at room temperature. For HCO_3^- 20 mL of the filtered extract is titrated against 0.02 N H_2SO_4 thrice to obtain concordant anion(s) concentration. Organic C and total N is estimated respectively by Walkley and Black wet oxidation method and Kjeldahl method.

2.4 Bio-chemical Analysis of Macrophytes

The macrophytic biomass of the emergent *M. aquaticum* (*M.a.*); the rooted floating *N. nucifera* (*N.n.*); the submerged *C. demersum* (*C.d.*) and the free floating *S. natans* (*S.n.*) were analysed for total chlorophyll content, biomass, productivity, specific growth rate, mineral composition (Ca, Cl, K, Mg, N, Na, P, S, Si) and the trace elements during their sprouting (sp.)-(March to May), peak growth (pg.)-(June to October) and senescence (sn.)-(November to February) phases. For plant and sediment samples various parametric evaluations are performed according to the protocol [23], [24] and [25]. The multi-elemental quantification of Ca, Cl, K, Mg, Na, P, S and Si in sediments and plant biomass is performed using Wavelength Dispersive X-ray Fluorescence Spectrometer (WD-XRF) at Central Instrumentation Facility (CIF), Pondicherry University. However, the trace elements in all samples were determined at Sophisticated Analytical Instrument Facility (SAIF) IITM, Chennai using ICP-OES (Perkin Elmer Optima 5300 DV) instrument.

2.5 Statistical Methods

The descriptive and illustrative statistical analysis including Mean, Standard Deviation, Range, Correlation, Regression, one-way ANOVA, Cluster Analysis and Principal Components Analysis (PCA) is performed using statistical software of MS-Excel 2010, Origin 8.5, PAST 3 and SPSS 19.

3 Results

The results of the study establish interdependent and interactive compartmentalization of inorganic elements in lacustrine components of water, sediment and macrophytes on space and time scales.

3.1 Surface Waters

The inter-annual monthly quantitative analysis (average values) of water quality parameters present a vivid inconsistency. The mean surface water temperatures recorded for the five study sites from March 2011 to February 2013 reflect a distinct contrast with higher values in summers and a declining trend towards the winters as a season specific feature. It ranged between 5 to 30°C with averages of 18.6±7.3, 17.9±7.2, 18.1±7.4, 18.1±7.0, 18.5±7.0 recorded at the five sites. The temperature variance between sites ($F = .013$ and $p > 0.05$) is insignificant, however, seasonal variation is significant ($F = 1.39$ and $p < 0.05$). The statistical insignificance in recorded epilimnion temperature variance as well as significant temporal discrepancy discloses similar geo-climatic setting of the lakes. pH fluctuated between 7.2 and 8.8 symbolizing alkaline waters with comparatively higher summer peaks established by macrophytic productivity. The site-wise averages are 8.4±0.32, 8.5±0.28, 8.2±0.43, 8.1±0.43 and 8.0±0.41. ANOVA for pH has significance between lakes ($F = 7.807$ and $p < 0.05$) and between seasons ($F = 9.782$ and $p < 0.05$) as well. Although the entire recorded pH profile (> 7) reflects buffered waters but has prominent variance both site- and time-wise suggesting position and period specific photosynthetic activity that withdraws HCO_3^- thereby elevating diurnal pH. Conductivity exhibited a fall towards the spring and summer months presumably due to dilution and ion uptake by flora present in the water-column. The conductivity range observed has a minimum value of 100 $\mu\text{S}/\text{Cm}$ and a maximum of 565 $\mu\text{S}/\text{Cm}$ besides having average site values of 353.9±57.46, 306.4±48.88, 294.7±113.83, 266.8±100.87 and 271.3±87.60. Seasonal ANOVA of conductivity is more significant ($F = 7.829$ and $p < 0.05$) than lake comparisons

($F = 4.526$ and $p < 0.05$). The diminishing conductivity at improved temperature conditions suggest higher bioaccumulation rate of nutrient ions like HCO_3^- , Ca, Mg, Na, K, P. The examination of the correlation matrix verily suggests temperature increments elevating pH but diminishing conductivity, HCO_3^- , Ca, Mg, Na, K and P significantly. However, Cl and $\text{NO}_3\text{-N}$ remain unaffected due to temperature and pH ordeal. The water characteristics of bicarbonates, chloride, calcium, magnesium, sodium, potassium, nitrate nitrogen and total phosphorus although differ in being as major ($> 5 \text{ mg/L}$) or minor (0.01 to 10 mg/L) elements but show almost an identical scenario of retreat during peak growth phases of macrophytes and again pickup in post senescence period. The PCA results reveal that variables of conductivity, K, Na, Mg, Ca, P, HCO_3^- , $\text{NO}_3\text{-N}$ and Cl contribute more on PC1 (60.27%) whereas temperature and pH predominate on PC2 (28.73%). An equation with the world average stream content (mgL^{-1}) of recorded Ca (> 15), Mg (> 4), K (> 2.3) and Na (> 6.3) acclaim pronounced human inputs from common uses like fertilizers, plaster, pigments, lime, alloys, pharmaceuticals, batteries, food additives, glass, baking powder, soft drinks, electroplating, caustic soda, water treatment chemicals, etc. The HCO_3^- concentration ranges from 60 to 218 mg/L and at the selected sites average to 126.2 ± 42.57 , 121.8 ± 41.77 , 105.2 ± 24.93 , 105.0 ± 25.89 and 105.7 ± 27.95 respectively. Bicarbonates register significant seasonal variation ($F = 35.431$ and $p < 0.05$) only. The waters are well buffered having insignificant site variation but there is gradual increase in summers due to higher proton uptake during peak photosynthetic activity thereby shifting CO_2 and HCO_3^- equilibrium. Contrarily, lower pH during winter supposedly on account of excess carbonic acid generation is due to higher [Respiration : Photosynthesis] proportion. Similarly, The ANOVA classifies other variables as significantly fluctuating between sites ($p < 0.05$) except Ca but insignificantly varying between seasons ($p > 0.05$) except Ca, Mg and K. The respective temporal range and site averages for these elements are presented in Table 1. ANOVA reveals significant seasonal variance in temperature, pH, conductivity and BIC compared to significant site contrasts for Cl, Mg, Na, K, $\text{NO}_3\text{-N}$ and P. The bicarbonates, chloride, calcium, magnesium, sodium, potassium, nitrate nitrogen and total phosphorus show almost an identical scenario of retreat during peak growth phases of macrophytes and again pickup in post senescence period. The comparative results exhibit least inter se difference in temperature, pH, Na, K, bicarbonates, Cl, Mg but larger fluctuations in other parameters of $\text{NO}_3\text{-N}$, P, Ca and conductivity. The similarity index Fig.2 of the sites in form of a dendrogram derived from hierarchical cluster analysis of water variables suggests a site specific peculiarity dependent on micro local conditions and perturbation types existing in the lakes. The correlation matrix provided in Table 2 shows Pearsons correlation coefficients and statistical significance with positive and negative significantly related water quality variables in bold asterisks. The regression analysis of water variables along with regression coefficients given in Fig.3 portrays vivid intra-aqueous interdependence by describing line of best-fit (R^2). The correlation matrix and regression analysis of water variables reflected in Table 2 and Fig.3 deduce a strong positive correlation between temperature and pH ($p < 0.01$, $R^2 = 0.8617$); temperature and conductivity ($p < 0.01$, $R^2 = 0.8419$); temperature or conductivity with other ions and intra-aqueous ionic composition ($p < 0.01$, $R^2 > 0.5$). The supportive understanding of water variable interaction is offered by significant coefficient of determination (R^2) alongside simple least square regression lines for pH versus temperature (0.8), conductivity versus pH (0.9), conductivity versus temperature (0.8) and BIC versus temperature (0.5). The graphical presentation in Fig.4 illustrated average trace element concentration in lake waters and their seasonal changes. The cationic content revealed a progression order of $\text{Ca} > \text{Mg} > \text{Na} > \text{K}$ all along with anionic predominance of HCO_3^- and Cl. The overall total P limit surpassed critical eutrophic index ($\leq 0.05 \text{ mgL}^{-1}$) but nitrate N persisted underneath it ($\leq 0.5 \text{ mgL}^{-1}$) mostly. $\text{NO}_3\text{-N}$ shows the only negative association with HCO_3^- concluding eutrophication steers on C assimilation for it serves as the main inorganic N source. The close covariance of P with Ca, Mg, Na and K depict their analogous limited absorbance constraint by the biota. The trace elements do not follow any definite seasonal variability. But ANOVA between sites confirm significance for Al ($F = 49.14$ and $p < 0.05$), Cr ($F = 3.854$ and $p < 0.05$), Cu ($F = 28.64$ and $p < 0.05$) and Ni ($F = 7.05$ and $p < 0.05$). Besides as per Table 3, Al shows significant correlations with Co, Cu and Ni while as association of As is significant with Cd and Ni. While comparing the observed average epilimnion trace element concentration (μgL^{-1}) with the average global stream abundance in absence of any baseline data, a vivid skewed hike for As ($\sim 10 > 2$), Cd ($\sim 3 > 1$), Cr ($\sim 40 > 1$), Co ($\sim 10 > 0.2$), Cu ($\sim 130 > 10$), Fe ($\sim 5600 > 700$), Pb ($\sim 90 > 3$), Mn ($\sim 600 > 7$), Ni ($\sim 30 > 1$), Se ($\sim 9 > 0.2$), Sn ($\sim 190 > 0.1$) and Zn ($\sim 18520 > 20$) proclaim their anthropogenic geneses. The exception of Fe ($> 5 \text{ mgL}^{-1}$) and Zn ($> 2 \text{ mgL}^{-1}$) enhance the maximal concentrations in

irrigation recommendations. Still, Zn has lesser toxicity in organic soils at pH > 6 and Fe contributes towards acidification and reduced P availability. Al (a non-priority pollutant) at pH 6.5 to 9 is well below the freshwater Criterion Continuous Concentration (CCC) chronic exposure limit of $\geq 87 \mu\text{g/L}$. Fe exceeds CCC at $> 1000 \mu\text{g/L}$ too. But all the priority pollutants of As, Cd, Cr, Cu, Pb, Ni and Se remained below chronic level of CCC except Zn.

Table 1. Temporal range and site averages for elements in surface waters

ELEMENT	RANGE	A1	A2	D1	D2	D3
Cl (mg/L)	11.6- 28.9	20.1±3.40	17.0±1.92	17.1±3.06	21.4±3.07	16.3±1.44
Ca (mg/L)	38 - 102	88.1±8.39	86.2±7.56	75.1±17.34	75.6±17.09	71.4±16.2
Mg (mg/L)	16.8 - 26	22.6±2.26	19.4±1.28	20.6±1.36	22.4±1.59	19.8±1.05
Na (mg/L)	7.5 - 13.4	10.6±1.23	9.4±0.66	8.8±0.47	10.0±0.72	8.9±0.52
K (mg/L)	2.6 - 5.3	4.4±0.70	3.9±0.52	3.9±0.45	4.4±0.49	3.6±0.42
NO ₃ -N (µg/L)	220 - 557	390.3±42.6	351.3±45.7	261.4±25.1	338.4±57.4	279.3±23.5
P (µg/L)	110 - 290	238.17±20.36	197.96±20.68	150.67±13.70	150.67±18.46	143.42±12.66

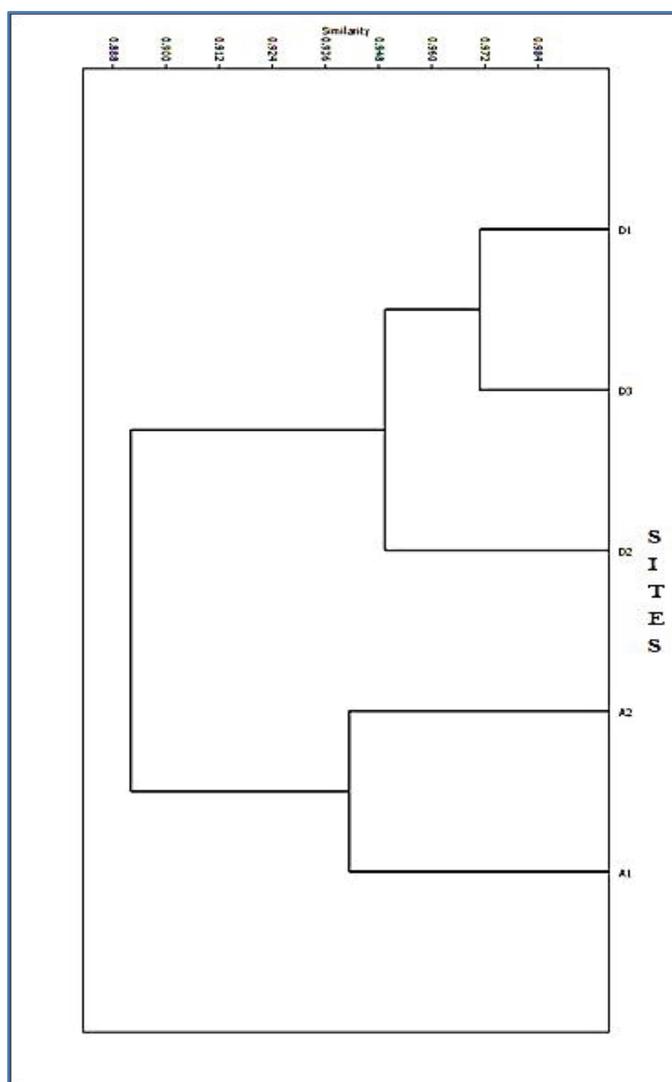


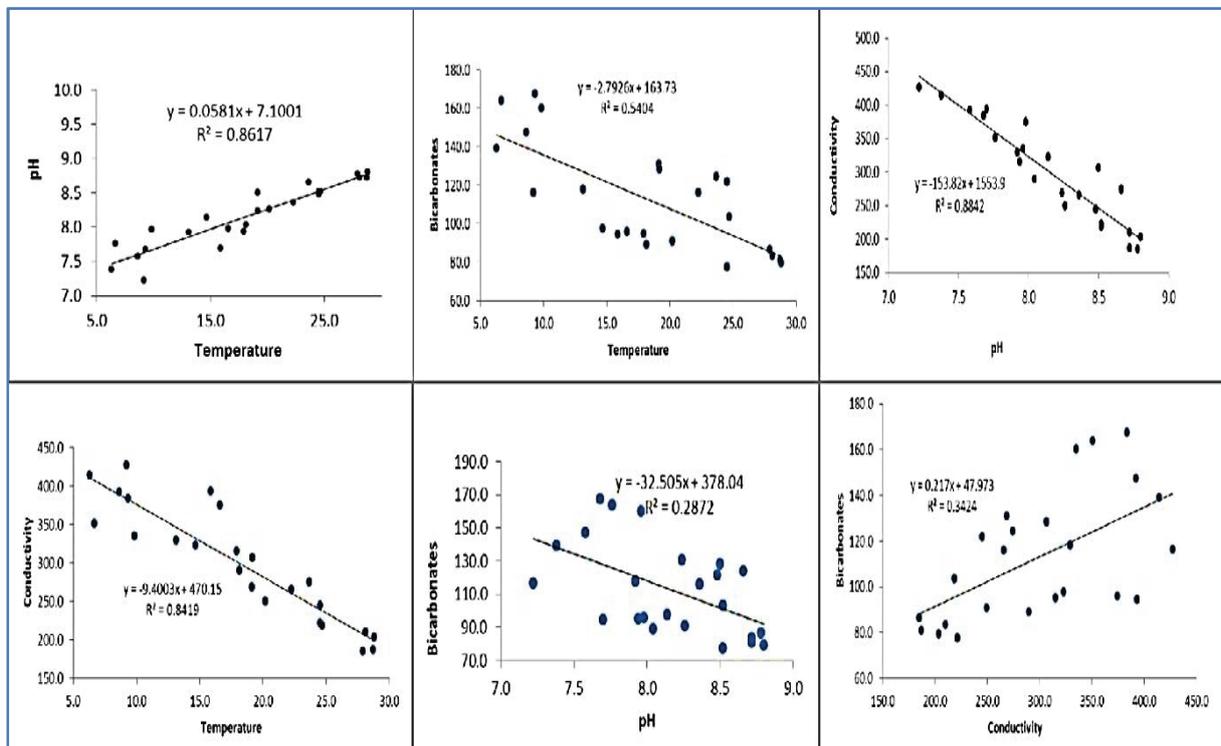
Figure 2. Dendrogram (Similarity Index) derived from hierarchical cluster analysis of water variables.

Table 2. Significance levels of water variables at 0.01 and 0.05 (2-tailed) correlations

		T	pH	Conductivity	Bicarbonates	Cl	Ca	Mg	Na	K	Nitrate-N	P
T	Pearson Correlation	1										
	Sig. (2-tailed)											
pH	Pearson Correlation	.918**	1									
	Sig. (2-tailed)	.000										
Conductivity	Pearson Correlation	-.917**	-.932**	1								
	Sig. (2-tailed)	.000	.000									
Bicarbonates	Pearson Correlation	-.736**	-.507*	.585**	1							
	Sig. (2-tailed)	.000	.012	.003								
Cl	Pearson Correlation	.266	.159	-.308	-.228	1						
	Sig. (2-tailed)	.209	.459	.143	.285							
Ca	Pearson Correlation	-.531**	-.568**	.617**	.222	-.114	1					
	Sig. (2-tailed)	.008	.004	.001	.296	.595						
Mg	Pearson Correlation	-.803**	-.715**	.800**	.613**	-.307	.329	1				
	Sig. (2-tailed)	.000	.000	.000	.001	.144	.116					
Na	Pearson Correlation	-.641**	-.672**	.641**	.299	.173	.532**	.590**	1			
	Sig. (2-tailed)	.001	.000	.001	.156	.420	.007	.002				
K	Pearson Correlation	-.661**	-.726**	.709**	.249	.258	.677**	.497*	.813**	1		
	Sig. (2-tailed)	.000	.000	.000	.240	.223	.000	.014	.000			
Nitrate- N	Pearson Correlation	.356	.382	-.312	-.408*	.295	.210	-.124	.083	.073	1	
	Sig. (2-tailed)	.088	.065	.138	.048	.162	.324	.563	.700	.736		
P	Pearson Correlation	-.485*	-.530**	.531**	.187	.343	.614**	.414*	.745**	.784**	.378	1
	Sig. (2-tailed)	.016	.008	.008	.381	.101	.001	.044	.000	.000	.069	

** Correlation is significant at 0.01 level (2-tailed)

* Correlation is significant at 0.05 level (2-tailed)

**Figure 3.** Line of best-fit (R^2) describing intra-aqueous interdependence of water variables.

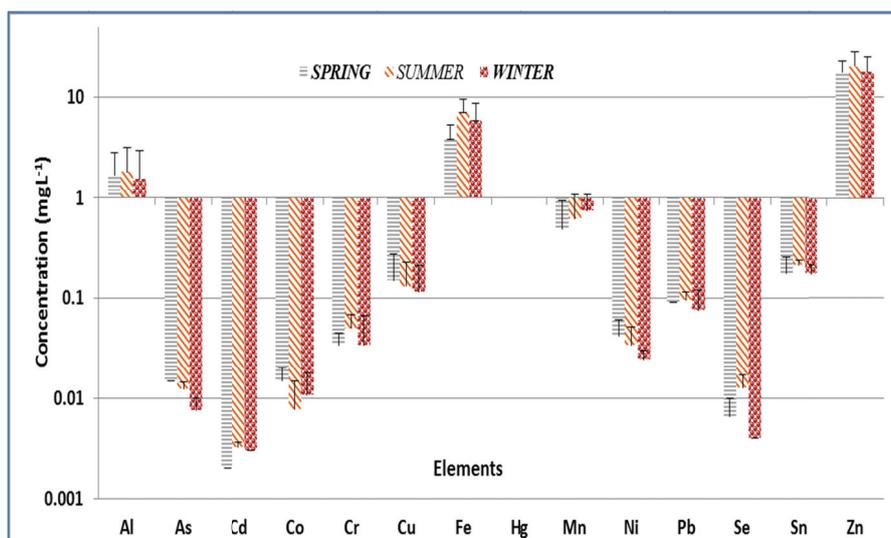


Figure 4. Average trace-metal concentrations and seasonal variation in micro and trace elements in lake waters

Table 3. Significance levels of trace elements in epilimnion at 0.01 and 0.05 (2-tailed) correlations

		Al	As	Cd	Co	Cr	Cu	Fe	Hg	Mn	Ni	Pb	Se	Sn	Zn
Al	Pearson Correlation	1													
	Sig. (2-tailed)														
As	Pearson Correlation	.350	1												
	Sig. (2-tailed)	.201													
Cd	Pearson Correlation	.152	.898**	1											
	Sig. (2-tailed)	.589	.000												
Co	Pearson Correlation	.652**	.488	.285	1										
	Sig. (2-tailed)	.008	.065	.303											
Cr	Pearson Correlation	.863**	.325	.200	.412	1									
	Sig. (2-tailed)	.000	.237	.475	.127										
Cu	Pearson Correlation	.870**	.448	.227	.769**	.766**	1								
	Sig. (2-tailed)	.000	.094	.416	.001	.001									
Fe	Pearson Correlation	-.213	.399	.610*	.015	-.134	-.047	1							
	Sig. (2-tailed)	.447	.141	.016	.959	.635	.868								
Hg	Pearson Correlation	. ^a	1												
	Sig. (2-tailed)														
Mn	Pearson Correlation	.323	-.135	-.176	.163	.339	.139	-.127	. ^a	1					
	Sig. (2-tailed)	.241	.632	.531	.561	.216	.621	.652	.						
Ni	Pearson Correlation	.543*	.661**	.485	.626*	.455	.520*	.019	. ^a	.289	1				
	Sig. (2-tailed)	.036	.007	.067	.013	.089	.047	.945	.	.297					
Pb	Pearson Correlation	.115	.492	.671**	-.019	.292	.203	.555*	. ^a	-.059	.264	1			
	Sig. (2-tailed)	.682	.062	.006	.945	.292	.468	.032	.	.834	.341				
Se	Pearson Correlation	-.180	.417	.473	-.353	.034	-.124	.245	. ^a	-.427	-.074	.373	1		
	Sig. (2-tailed)	.521	.122	.075	.197	.905	.659	.380	.	.112	.795	.171			
Sn	Pearson Correlation	.276	.409	.510	.166	.457	.219	.491	. ^a	.399	.502	.433	.073	1	
	Sig. (2-tailed)	.319	.130	.052	.553	.087	.432	.063	.	.140	.057	.107	.795		
Zn	Pearson Correlation	.088	.099	.291	-.125	.167	.043	.224	. ^a	.209	.315	.591*	.149	.283	1
	Sig. (2-tailed)	.755	.725	.292	.658	.552	.880	.422	.	.455	.253	.020	.597	.306	

** Correlation is significant at 0.01 level (2-tailed)

a Cannot be computed because at least one of the variables is constant (BDL)

* Correlation is significant at 0.05 level (2-tailed)

3.2 Surface Sediments

The observed range of H-ion concentration in surface sediments is 6.16 to 8.35 and the respective mean site values include 7.22 ± 0.60 , 7.30 ± 0.64 , 7.58 ± 0.42 , 7.20 ± 0.58 , 7.49 ± 0.56 . The sediment pH generally remained on the basic side but slight acidic nature is noticed during late summer probably due to temperature determined decomposition of organic matter releasing acidic by-products. Alkaline pH records of the sediment in general determine the catchment calcite predominance whereas slight summer diminution is temperature improved OM decomposition effect. Sediment conductivity ranged between 234 and $498\mu\text{S}/\text{cm}$ at 25°C and the average inter-site comparison is as 395.4 ± 55.7 , 312.6 ± 47.2 , 293.3 ± 36.4 , 363.3 ± 50.9 and 323.3 ± 38.4 respectively. The sediment bicarbonate content has minima of 116 and maxima of 264 mg/Kg and varied between the sites as 187.4 ± 25.3 , 193.5 ± 37.2 , 169.8 ± 41.5 , 172.6 ± 41.9 and 189.6 ± 42.3 . Similarly, ranges and average site percentage dry-weight variations in sediment organic C, organic matter, total N and C/N ratio are summarized in Table 4. A temporal trend of the above seven sediment characteristics for contiguous eight seasons of two year study can be visualised as inconsequential whereas Ca-Si dominance is depicted throughout. Table 5 depicts significant correlation for conductivity with organic C and organic matter as the latter serves an established source for nutrient ions. Similar is the case with sediment organic C, organic matter and total N as being complementary to one another. The typical < 10 C/N implies autochthonous OM, besides low decomposition rates controlled by N budgeting and speciation via biochemical (enzyme and pH) fluctuations under alkaline conditions. Higher temperature and lower N:P ratio during summer develop the P internal loading process. This P recycling mediates via degradation (oxic), denitrification and sulphate reduction. Nonetheless, higher Al, Ca and Fe proportions in sediments inactivate P mobilization. The estimated percentage elemental composition (Fig. 5) of surface sediments on dry weight basis revealed the order $\text{Si} > \text{Ca} > \text{Mg} > \text{K} > \text{Na} > \text{P} > \text{S} > \text{Cl}$. Besides, the micro and trace element quantification (Fig.5) represent the descending series of $\text{Fe} > \text{Al} > \text{Zn} > \text{Mn} > \text{Cu} > \text{Cr} > \text{Ni} > \text{Co} > \text{As} > \text{Sn} > \text{Pb} > \text{Cd}$ while Hg and Se remained below detection level (BDL).

Table 4. Ranges and average site percentage dry-weights of sediment variables

SEDIMENT-VARIABLE	RANGE	A1	A2	D1	D2	D3
ORGANIC-CARBON (%)	6.7 to 15.21	11.05 ± 2.38	9.42 ± 1.58	8.67 ± 0.84	9.83 ± 1.27	7.85 ± 0.77
ORGANIC-MATTER (%)	11.5 to 26.22	19.035 ± 4.096	16.238 ± 2.726	14.940 ± 1.442	16.939 ± 2.178	13.526 ± 1.327
TOTAL-NITROGEN (%)	1.09 to 2.87	1.948 ± 0.332	1.828 ± 0.322	1.864 ± 0.306	1.900 ± 0.174	2.458 ± 0.322
C/N RATIO	2.55 to 7.53	5.710 ± 1.018	5.208 ± 0.720	4.789 ± 0.948	5.175 ± 0.554	3.229 ± 0.373

Table 5. Significance levels of sediment variables at 0.01 and 0.05 (2-tailed) correlations

		pH	Conductivity	Bicarbonate	OC	OM	TN	C/N ratio
pH	Pearson Correlation	1						
	Sig. (2-tailed)							
Conductivity	Pearson Correlation	-.350	1					
	Sig. (2-tailed)	.395						
Bicarbonate	Pearson Correlation	.720*	-.039	1				
	Sig. (2-tailed)	.044	.928					
OC	Pearson Correlation	.086	.853**	.357	1			
	Sig. (2-tailed)	.840	.007	.385				
OM	Pearson Correlation	.086	.854**	.356	1.000**	1		
	Sig. (2-tailed)	.840	.007	.386	.000			
TN	Pearson Correlation	.320	.678	.504	.860**	.860**	1	
	Sig. (2-tailed)	.439	.065	.203	.006	.006		
C/N ratio	Pearson Correlation	-.545	.279	-.275	.130	.129	-.352	1
	Sig. (2-tailed)	.163	.503	.510	.759	.760	.392	

* Correlation is significant at 0.05 level (2-tailed)

** Correlation is significant at 0.01 level (2-tailed)

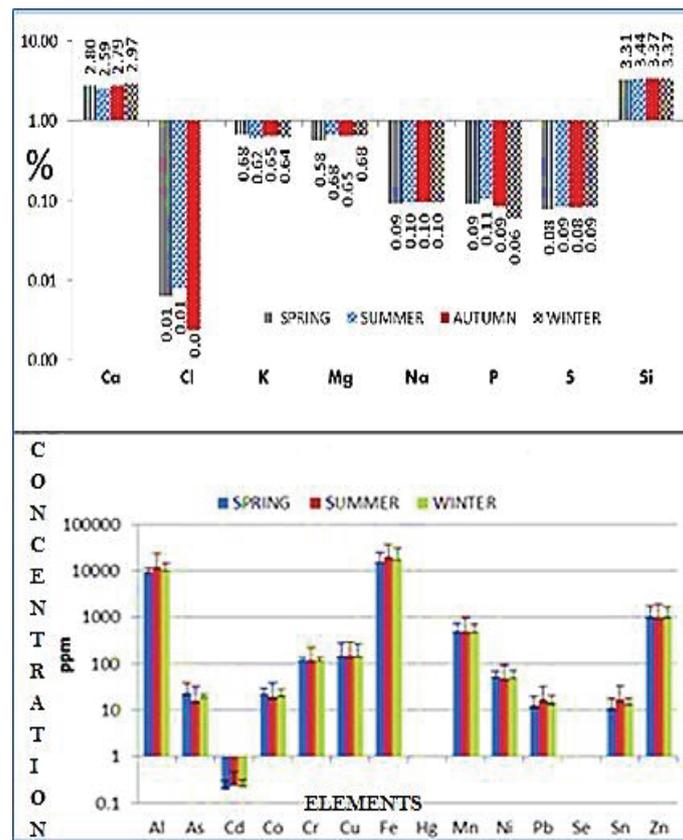


Figure 5. Mean percentage elemental composition of surface sediments.

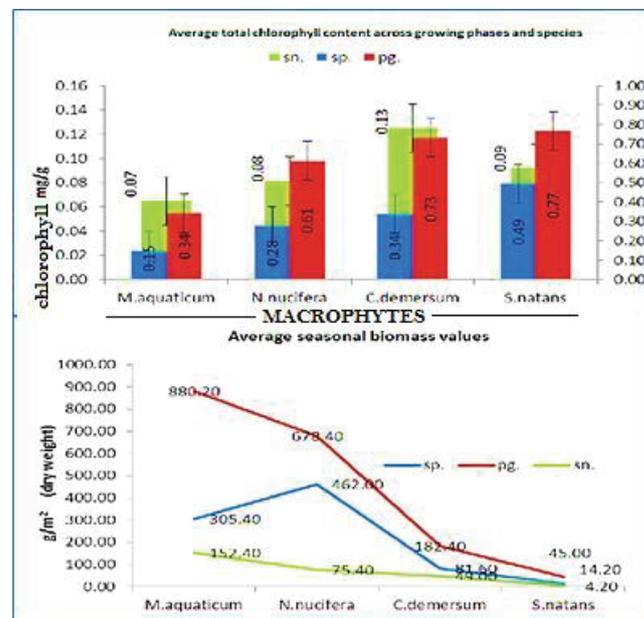


Figure 6. Total chlorophyll and dry-weight biomass of the experimental species.

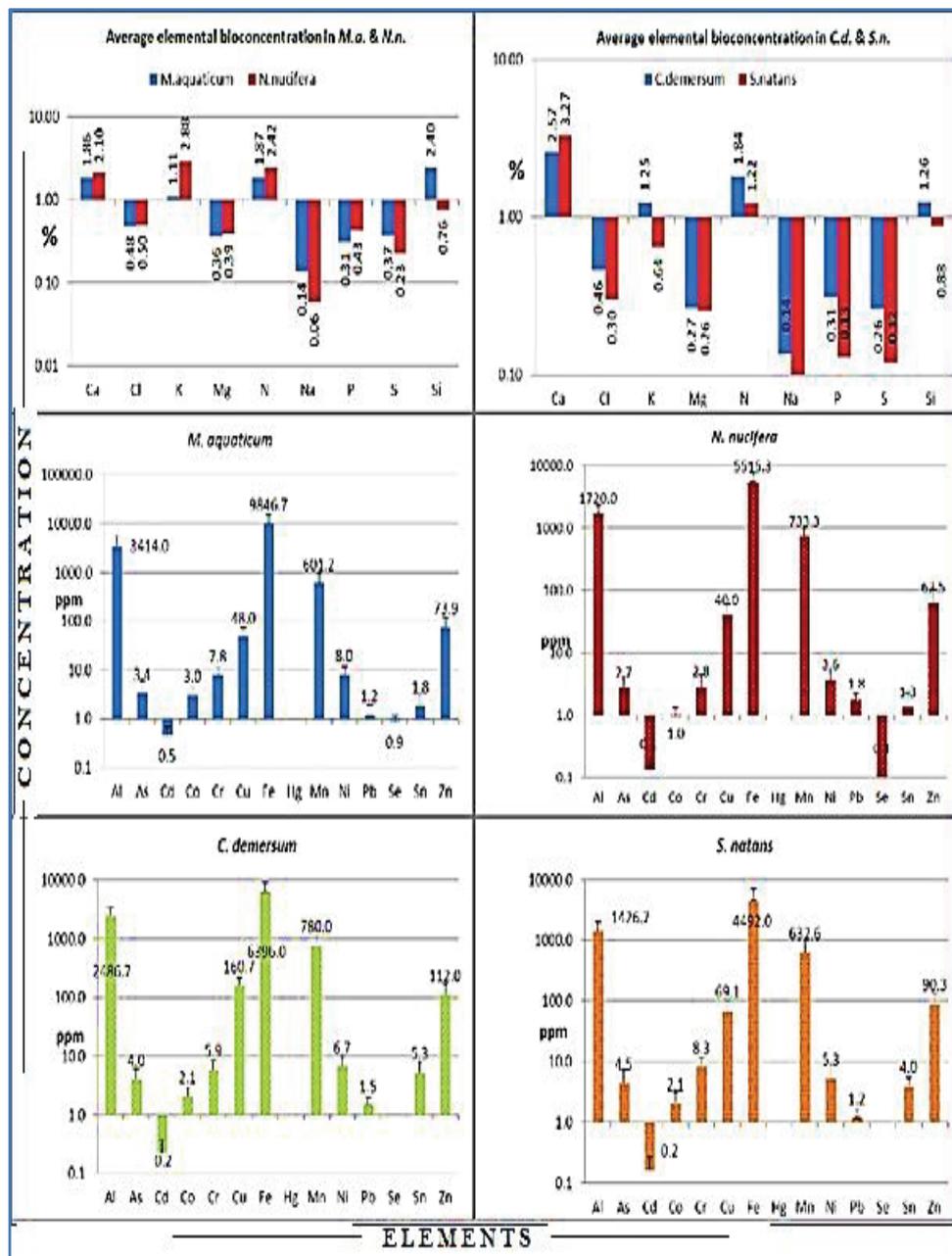


Figure 7. Bioaccumulation profile of nutrients and trace elements.

3.3 Macrophytes

The productive capacity of the investigated macrophytes calculated on the basis of total chlorophyll content and dry-weight biomass for each experimental species in the three major growth phases is presented in Fig.6. Further, the tissue concentration of nutrients and trace elements illustrated in Fig.7 differ species wise but well correlate with ambient water and sediment media. *M. aquaticum* accumulated 49% of the dry weight biomass annually whereas for *N. nucifera* amounted to 38%, *C. demersum* to 10% and *S. natans* contributed only 3%. The analysis of four dominant ubiquitous life forms of macrophytes signified marked variance in productivity pattern and specific growth rates of the order emergent *M. aquaticum* > rooted floating *N. nucifera* > submerged *C. demersum* > free floating *S. natans* well in agreement to their physiognomy and micro-habitat.

4 Discussion

The limnochemical statistics that evolved during the study retro expound definite and obvious development of physicochemical and biological configuration of the select sites in both lakes under multiple anthropogenic stress [38]. The chemical composition of water closely related to dual factors of concerned catchment and human intervention thereof [39]. The lakes studied deem to be classified under eutrophic category with considerable differences in bio-physicochemical milieu due to varied nature of operative anthropogenic stress [28]. But for the semi-drainage type hydrology flushing via regular outflow channels supplemented by ephemeral and sub-aqueous inputs is vital for their self-reclaimed homeostasis since the basin holding-time of water keeps low ($2^{1/2}$ months). A higher pH associated with elevated temperature and alkaline conditions is reported to favour co-precipitation of carbonate and phosphate due to rapid carbon assimilation from dissolved bicarbonates [26] and [40]. Despite a temporary inverse relationship of bicarbonate concentration and pH during peak-growth summer days, the enriched alkaline pH switch ascribes to equivalent carbonate dominance [1] and [27]. Conductivity describes the trophic status in terms of total nutritive ionic strength of the water column. Its decline in peak growth season implies ionic depletion by photoautotrophic consumption and biogenic calcification (precipitation of CaCO_3) commonly referred to as seasonal whiting or clouding of lakes [41]. The ionic strength (conductivity) measures vindicate superior eutrophy (solute richness) besides autotrophic assimilation and biocalcification contributing towards its summer falls [28] and [42]. The lime grade catchment characteristic contributed to Ca dominance phenomenon and hard water type classification [33] and [43]. K features base association with agricultural runoff and Cl to faunal organic pollution [44] and [45]. A range of human actions involving agriculture, farming, sewage, factories, etc. result in N species inflow [46]. The spring fed basins of both the lakes and preferential reduced NH_4^+ autotrophic assimilation instead of oxidised NO_3^- also greater concentration of nitrate – N in water [29] and [47]. Additionally, isolated geogenic N pockets are ascribed as evaporative deposits in phyllite, schist and carbonate bed rocks in Kashmir [33]. Forest surface runoff too contain $\leq 20 \text{ mgL}^{-1}$ nitrate as a natural source to lakes [48]. Summer decreases of the least abundant macronutrient (P) limits bio-productivity owing to its utilization and carbonate coprecipitation [49]. The amount of N and P forms is intimately related to biological productivity of aquatic ecosystems [47]. Higher temperatures favour bio-production and thereby elevate pH significantly. So, the factors of temperature and pH clearly state that improved growth environment for biological uptake and calcite co-precipitation of ions reduce their epilimnion concentrations in peak growth summer cycle. Such a systematics attributed to photosynthetically induced precipitation and utilization by autotrophs is concomitant to latest studies [50]. Anthropogenic effluents are established primary sources of traces in surface waters [30] and [51]. Pertinently, such proportions exceed maximum permissible limits for drinking water but are suitable as irrigation liquid ordinarily [31], [32] and [52]. The complexing phenomena of heavy metals with carbonate and BIC limit their direct or indirect repercussions [53]. Now, the possible source identification includes solid and liquid discards from an array of utilities namely alloys, storage batteries, pesticides, wood preservatives, electroplating, pigments, fertilizers, porcelain, glass, electrical wiring, roofing, utensils, piping, chemicals, paints, vapour lamps, mirror coatings, thermometers, catalysts, ceramics, fossil fuels, electronics, solder and so on [17].

The abiotic quiescent zone of sediments is suitable long-term indicator of lake environmental conditions [54] and [55]. Sediment nutrient constitution corresponds to catchment land use [56]. They provide settling space for water carried elements, both acquired and innate, in the form of ions, organic/inorganic- complexes and dissolved/suspended- matter associations [57]. Sediments reflect weaker source but stronger sink behaviour and capacity depicted via transformations of adsorption/desorption, mineralization/demineralization, bioassimilation and burial phenomenon [58]. Sediments' pH can be labelled as the principal factor enacting nutrient accessibility and movement [59]. The seasonal organic C changes in sediments aptly describe a pattern of temperature induced decrement during summers fulfilling the mineralization promotion [33] and [60]. The analysis of superficial sediment chemical characters reveals connections with trophic index [61]. During the course of the study most of the nutrients could be observed to follow similar tendency both in surface sediments and overlying water column ominous of systematic interchange [62]. A general gradual decline in their concentration till culmination of active macrophytic growth phase can be associated with active/passive bioaccumulation

or anoxic release from sediments [63]. Its possible reason could be NH_4^+ microbial preference instead of NO_3^- stimulating decomposition despite eutrophied state [34] and [64]. Improved N:P ratios parallel chronic nitrogen influx in P-limited lakes [65], however, N-limited lakes suffer greater eutrophication catastrophes [66]. Curbing the external N and P loads is effective in remediation but the sediment internal loading supplement compensates the loss [35] and [67]. Meanwhile sediment dredging is beneficial for internal nutrient deloading only after the external inputs are checked [68]. Trace metals bound to OM or Fe/Mn- oxides separate on their decomposition and reductive dissolution respectively near hypolimnion-sediment overlap [35], [36] and [69]. The lower [Ca : Al] values suggest exsitu human cause for PTE transport to the designated lacustrine sites [58]. Nonetheless, the mobility of PTE's is effectively curtailed by OM enriched sediments and calcite co-precipitation [70].

Macrophytes dominate mineral regulation between sediment and water [71]. They are capable to remove nutrients even at low loading rates [72]. Their adaptive advantage to lock up minerals from nutrient pool is helpful to check degradation primarily and restore mineral loading later on [47]. N and P are intimately related to biological productivity of aquatic ecosystems [37]. The productivity also lies in consonance with developmental stage, ambient nutrient medium and physiognomy of a species [74]. Higher temperature favour bio-production and elevate carbonate driven pH significantly [75]. The peak nutrient uptake and bioconcentration coincided with peak biomass in summer and autumn [37], [76] and [77]. Substantial OM productions by macrophytes contribute towards nutrient immobilization and provide OC requirement for denitrification while decaying [78]. The upgraded biomass configuration property establishes an additional nutrient and trace element interception feature of hyperaccumulation from the growth medium [79]. Consequently, macrophytes act as unselective nutrient sinks during their peak growth period conferring them rhizofiltration and phytofiltration attributes.

Conclusively, the study discovers that natural setting and homeostasis is influenced by multiple operational factors and pollution conditions. Each site offers a differing dynamics in the trophic ranking of $A1 > A2 > D2 > D3 > D1$ in relation to the type and magnitude of human interference. The structural parts of water, sediment and macrophytes are essential to dissipate the contaminants temporarily to a certain limit. The complications of mismanaged systems are grave as opposed to the managed ones. The study paves way to associate concentration versus speciation connections for diverse pollutants and their on-site phytofiltration options. Thus ecorestoration practices along with retaining optimization capacity can decelerate deterioration.

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